

Middle-range theories of land system change

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Meyfroidt, P.^{1,2*}, Roy Chowdhury, R.³, de Bremond, A.^{4,5}, Ellis, E.C.⁶, Erb, K.-H.⁷, Filatova, T.^{8,9},
Garrett, R.D.¹⁰, Grove, J.M.¹¹, Heinemann, A.^{5,12}, Kuemmerle, T.^{13,14}, Kull, C.A.¹⁵, Lambin, E.F.^{1,16,17},
Landon, Y.¹⁸, le Polain de Waroux, Y.¹⁹, Messerli, P.^{5,12}, Müller, D.^{13,14,20}, Nielsen, J. Ø.^{13,14}, Peterson,
G.D.²¹, Rodriguez García, V.¹, Schlüter, M.²¹, Turner II, B. L.²², Verburg, P.H.^{23,24}

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*: Corresponding author. patrick.meyfroidt@uclouvain.be

Université catholique de Louvain, Earth and Life Institute

Georges Lemaître Centre for Earth and Climate Research (TECLIM)

Place Pasteur 3, bte L4.03.08

1348 Louvain-la-Neuve

Belgium

1. Georges Lemaître Center for Earth and Climate Research, Earth and Life Institute, Université catholique de Louvain, 1348

Louvain-la-Neuve, Belgium

2. F.R.S.-FNRS, 1000 Brussels, Belgium

3. Graduate School of Geography, Clark University, Worcester, MA 01610-1477

4. Geographical Sciences Department, University of Maryland, College Park, MD, USA

5. Centre for Development and Environment (CDE), University of Bern, Bern, Switzerland.

6. Geography & Environmental Systems, University of Maryland, Baltimore County, 1000 Hilltop Circle, Baltimore, MD 21250

USA

7. Institute of Social Ecology (SEC), Department of Economics and Social Sciences (WiSo), University of Natural Resources and

Life Sciences (BOKU), Schottenfeldgasse 29, A-1070, Vienna, Austria

8. Department of Governance and Technology for Sustainability, University of Twente, P.O. Box 217, 7500 AE Enschede, The

Netherlands

- 39 9. School of Systems (SML), Faculty of Engineering and Information Technology, University of Technology Sydney, Sydney, NSW
40 2007, Australia
- 41 10. Department of Earth and Environment, Boston University, Boston, MA, USA
- 42 11. USDA Forest Service, Suite 350, 5523 Research Park Drive, Baltimore, MD, 21228, e:morgangrove@fs.fed.us
- 43 12. Institute of Geography, University of Bern, Bern, Switzerland
- 44 13. Geography Department, Humboldt-University Berlin, Unter den Linden 6, 10099 Berlin, Germany
- 45 14. Integrative Research Institute on Transformations of Human-Environment Systems (IRI THESys), Humboldt-University Berlin,
46 Unter den Linden 6, 10099 Berlin, Germany
- 47 15. Institute for Geography and Sustainability, University of Lausanne, 1015 Lausanne, Switzerland
- 48 16. School of Earth, Energy & Environmental Sciences, Stanford University, Stanford, CA, USA
- 49 17. Woods Institute for the Environment, Stanford University, Stanford, CA, USA
- 50 18. School of Public and Environmental Affairs, Indiana University, Bloomington, IN 47405, US.
- 51 19. Institute for the Study of International Development and Department of Geography, McGill University, Montreal, QC, Canada.
- 52 20. Leibniz Institute of Agricultural Development in Transition Economies (IAMO), Theodor-Lieser-Str. 2, 06120 Halle (Saale),
53 Germany
- 54 21. Stockholm Resilience Centre, Stockholm University, 10691 Stockholm, Sweden.
- 55 22. School of Geographical Sciences and Urban Planning & School of Sustainability, Arizona State University, PO Box 875302,
56 Tempe, Arizona, 85287-5302, USA
- 57 23. Environmental Geography Group, Institute for Environmental Studies (IVM), Vrije Universiteit Amsterdam, Amsterdam 1081
58 HV, The Netherlands
- 59 24. Swiss Federal Institute for Forest, Snow and Landscape Research (WSL), Birmensdorf, Switzerland

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81 Abstract

82 Land system changes generate many sustainability challenges. Identifying more sustainable land-
83 use alternatives requires solid theoretical foundations on the causes of land-use/cover changes. Land
84 system science is a maturing field that has produced a wealth of methodological innovations and
85 empirical observations on land-cover and land-use change, from patterns and processes to causes.
86 We take stock of this knowledge by reviewing and synthesizing the theories that explain the causal
87 mechanisms of land-use change, including systemic linkages between distant land-use changes,
88 with a focus on agriculture and forestry processes. We first review theories explaining changes in
89 *land-use extent*, such as agricultural expansion, deforestation, frontier development, and land
90 abandonment, and changes in *land-use intensity*, such as agricultural intensification and
91 disintensification. We then synthesize theories of higher-level land system change processes,
92 focusing on: (i) *land-use spillovers*, including land sparing and rebound effects with intensification,
93 leakage, indirect land-use change, and land-use displacement, and (ii) *land-use transitions*, defined
94 as structural non-linear changes in land systems, including forest transitions. Theories focusing on
95 the causes of land system changes span theoretically and epistemologically disparate knowledge
96 domains and build from deductive, abductive, and inductive approaches. A grand, integrated theory
97 of land system change remains elusive. Yet, we show that middle-range theories – defined here as
98 contextual generalizations that describe chains of causal mechanisms explaining a well-bounded
99 range of phenomena, as well as the conditions that trigger, enable, or prevent these causal chains –,
100 provide a path towards generalized knowledge of land systems. This knowledge can support
101 progress towards sustainable social-ecological systems.

102

103 **Keywords:** human-environment systems; box and arrow framework; indirect land-use change;
104 land-use intensification; deforestation; land-use spillover; urban dynamics.

105

106 1. Introduction

107 Change in land use—the purposes and activities through which people interact with land and
108 terrestrial ecosystems— is a key process of global environmental change. Understanding land-use
109 change is central for designing strategies to address sustainability challenges, including climate

110 change, food security, energy transition, and biodiversity loss. Land systems constitute complex,
111 adaptive social-ecological systems (Berkes et al. 1998) shaped by interactions between (i) the
112 different actors and demands that act upon land, (ii) the technologies, institutions, and cultural
113 practices through which societies shape land use, and (iii) feedbacks between land use and
114 environmental dynamics (MA 2003, Verburg et al. 2015). Elementary events of land-use changes
115 that take place at the plot-level over short time periods, such as deforestation or substitution of one
116 crop by another, correspond to changes in the extent and/or intensity of land use. These elementary
117 building blocks combine to form complex, structural processes taking place over broader extents
118 (landscapes, regions, and across countries) and longer time scales, including non-linear transitions
119 (Lambin and Meyfroidt 2010) and spatial reorganization of land uses (Rey Benayas et al. 2007,
120 Kastner et al. 2014, Queiroz et al. 2014, Levers et al. 2018).

121 Land system science is a maturing field that has produced a wealth of methodological innovations
122 and empirical observations (Lambin et al. 2006, Turner et al. 2007, Verburg et al. 2015). It focuses
123 on monitoring and describing patterns of land-cover change, explaining drivers of land-use change,
124 and understanding linkages between these two. These advances have relied on deductive approaches
125 based on disciplinary frameworks (e.g., neo-classical economics or political ecology), abductive
126 reasoning (i.e., starting from outcomes and retracing these to their likely causes), syntheses based
127 on systematic reviews and meta-analyses of drivers and impacts of land system change (Magliocca
128 et al. 2015, van Vliet et al. 2015), and “box and arrows” conceptual frameworks. The development
129 of land system theories has been lagging due to: (i) a focus on local case studies, favoring *ad hoc*
130 interpretations based on contingent factors; (ii) an emphasis on methodological developments
131 involving improvements in remote sensing and other geospatial analyses; and (iii) the
132 interdisciplinary nature of land system science, which has led to the borrowing of theories from
133 related disciplines including geography, landscape ecology, economics, and anthropology
134 (Meyfroidt 2015, 2016).

135 Lambin et al. (2001) challenged simplistic notions about the causes of land-use and land-cover
136 change, highlighting complex interactions, multi-causality, and the contextual character of land
137 system processes. Here, we argue that land system dynamics can be apprehended through
138 theoretical generalizations that transcend the place-based specificity of cases, without ignoring their
139 complexity. We consider that theoretical formalization can further the development of: (i) testable
140 hypotheses; (ii) process-based models simulating complex interactions; and (iii) credible knowledge
141 that informs policy and decision-making beyond specific places while remaining sensitive to
142 context. Theories of land systems advance our understanding of the dynamics of social-ecological
143 systems and foster dialogue with other human-environmental sciences.

144 Here, we take stock of land system science knowledge generated over the last decades, focusing on
145 theories explaining the causes of land-use change and their systemic linkages across places. We
146 focus on middle-range theories, defined as contextual generalizations presenting causal
147 explanations of delimited aspects of reality—events or phenomena (Merton 1968, full definition in
148 Section 2). This stands in contrast to both high-level, unified theories, as well as explanations
149 relying on the singularity of a specific case. While our focus is not on theories relating land-use
150 change to its environmental and human impacts, we account for feedback mechanisms that alter the
151 dynamics of land use. We thus only touch lightly on the normative aspects of land system change.
152 We concentrate on processes in agriculture and forestry, but many theories discussed here have been
153 used for other dynamics, such as urban land uses.

154 Our objective is to articulate how middle-range theories can contribute to understanding land
155 system change by:

- 156 (i) Reviewing the different theories explaining changes in land-use extent and intensity, and
- 157 (ii) Synthesizing them into middle-range theories of higher-level processes of land system
158 changes, focusing on land-use spillovers and land-use transitions as non-linear, structural
159 changes.

160 Section 2 discusses the role of middle-range theories in relation to frameworks, models, and
161 typologies. Section 3 reviews theories of land-use expansion and intensification. Sections 4 and 5
162 build on these theories to synthesize middle-range theories on structural changes in land systems.
163 We then discuss further theory development on land systems as social-ecological systems.

164

165 2. Theories, frameworks, models, and typologies

166 Different epistemologies have distinct visions of what a “theory” is. Here, a *theory* is defined as a
167 general explanation or stylized facts about events, phenomena, or their attributes (e.g., spatial or
168 temporal patterns), based on a set of factors and their causal relations. The term “*middle-range*
169 *theory*”, originating from social sciences, describes a process developing from observations and
170 analyses of a specific event or phenomenon, building towards explanations of sets of similar
171 phenomena, which can be progressively expanded to other phenomena presenting similar
172 characteristics or linked to other mechanisms present in other theories (Merton 1968). Here, we
173 define middle-range theories as *contextual generalizations that describe chains of causal*
174 *mechanisms explaining a well-bounded range of phenomena, as well as the conditions that trigger,*
175 *enable, or prevent these causal chains* (Meyfroidt 2016). Middle-range theories seek to balance
176 generality, realism, and precision across the breadth of explanatory factors mobilized, to reach a

177 middle ground between ad hoc explanations of singular cases and “grand”, universal systems
178 theories that explain all features in a stylized way (Levins 1966, Hedström and Udehn 2009,
179 Hedström and Ylikoski, 2010). In contrast with grand theories, which are posited to apply to a very
180 wide range of phenomena, middle-range theories tend to have a narrower focus and application and
181 should be explicit about the processes it aims to explain and the limits of its reach. Over time,
182 middle-range theories can expand their reach or be combined with each other, as the underlying
183 mechanisms that join them are better understood. Multiple disciplinary and interdisciplinary
184 middle-range theories have been proposed to explain land system changes (SI Appendix A, see
185 Sections 3-4-5).

186 Middle-range theories can be distinguished from other generalization approaches including
187 conceptual frameworks, models, and typologies. *Frameworks* are a collection of concepts
188 considered as relevant for analyzing a phenomenon, which constitute lenses for looking at reality
189 and boundary objects for inter- and transdisciplinary communication (McGinnis 2011). They
190 provide checklists of variables and components to include in theories, and indicate the assumed
191 structural relations between these building blocks. In contrast with theories, these relations are
192 neither depicted functionally nor their strength hypothesized under different sets of conditions.
193 Prominent frameworks in human-environmental science are the frameworks on proximate causes
194 and underlying drivers of environmental change (Geist and Lambin 2002, MA 2003), the social-
195 ecological systems framework depicting factors of sustainable self-organization of resource-use
196 systems (Ostrom 2009), ecosystem services frameworks linking human well-being and ecosystems
197 (Daily et al. 2009; MA 2003; Fisher et al. 2013, van Zanten et al. 2014), the telecoupling framework
198 on linkages between distant social-ecological systems (Liu et al. 2013, Friis et al. 2016), and others
199 (SI Appendix B). Examples of frameworks specific to land system science include those on major
200 land system components (GLP, 2005), distinction between land-use, land-cover and land-
201 management changes (Pongratz et al. 2018), and land-use intensity (Erb et al. 2013, Kuemmerle et
202 al. 2013). Many social science theories are somewhere between middle-range theories and
203 frameworks.

204 Frameworks and theories provide bases for constructing dynamic *models* aiming to replicate and
205 enhance system understanding by formalizing and exploring the relations between different
206 variables and their outcomes (National Research Council, 2014, Verburg et al., 2016). Process-
207 based models can rely on theories to inform their assumptions on the structure and type of relations
208 between variables. Models can play an important role in the development and testing of theories,
209 particularly to identify mechanisms and their effects under certain conditions, and to quantify the
210 relations between variables. Constructing a model can be akin to building a theory, by selecting

211 variables, generating hypotheses on their relations, and assessing their influence on outcomes. As
212 social-ecological systems are complex and adaptive, their dynamics are influenced by bottom-up
213 (emergent) and top-down (constraining/enabling) processes and structures. Top-down and bottom-
214 up mechanistic theories can be validated by implementing them in a process-based model, such as
215 agent-based models.

216 Another generalization approach involves the identification of *typologies*, also referred to as
217 syndromes or archetypes, i.e., recurring patterns or combinations of variables, processes, actors,
218 situations, or outcomes (Schellnhuber et al. 1997, Oberlack et al. 2016, Valbuena et al. 2008, Levers
219 et al. 2018). Typologies can be derived inductively by identifying commonalities within a set of
220 cases, for example using qualitative-comparative analysis (QCA) or other configurational
221 approaches, or deductively via the theoretical identification of key variables that create a
222 typological space. Typologies often lack causal relations, but can be used to build “typological
223 theories” (George and Bennett 2005).

224

225 3. Theories of land-use expansion and intensification

226 3.1 Land-use expansion and intensification

227 The increasing global demands that human societies place on land, including for production of
228 goods, nature protection, and ecosystem services, require changes in extent (expansion or
229 contraction) and intensity (intensification or disintensification¹) of land uses. Land-use expansion
230 occurs into unconverted areas (“wildlands”) or over land that is already converted to anthropogenic
231 land cover, such as cropland expanding over pastures as often observed in South America
232 (Baumann et al. 2017). Land-use intensification refers to practices that increase land productivity by
233 (i) increasing inputs per land unit (e.g., labor and capital-based inputs, or technology) or the
234 temporal frequency of land use (e.g., multiple harvests), (ii) increasing output per land unit (i.e.,
235 yields), and/or (iii) altering ecosystem properties, as with tree species homogenization in intensive
236 forestry (Erb et al. 2013, Kuemmerle et al. 2013). Considering multiple and growing demands
237 (Haberl et al. 2014, Haberl 2015), intensification is often seen as a path for sustainability, to lessen
238 competition for productive land and mitigate trade-offs such as between food security and
239 environmental conservation (see Section 4.1). However, intensification can produce multiple
240 undesirable environmental and social impacts, such as increasing capital costs to impoverished
241 smallholders (Luyssaert et al. 2014, Kremen 2015, Gossner et al. 2016, Erb et al. 2016a, 2017). The

1 “Extensification” is sometimes also used for disintensification, mainly in the European context, while the same term
is frequently used for “expansion” in the North American context. We therefore mainly use disintensification here.

3 Mirroring the multidimensional nature of intensification, disintensification can encompass various realities.

242 functions assigned to land thus inherently result from social dynamics and conflicting purposes,
243 interacting with biophysical factors. Land-use diversification, or an increase in multifunctionality to
244 produce different goods and ecosystem services on the same land, can be decomposed into the
245 intensification of some land uses and disintensification of others. This distinction is often employed
246 strategically to take advantage of land-use synergies to increase resilience or maintain a basket of
247 outputs per land unit while reducing non-land inputs (Fischer et al. 2017). Expansion and
248 intensification can co-occur, for example intensification through the expansion of a more intensive
249 land use over a less intensive one (Baumann et al. 2017, Meyfroidt et al. 2014).

250 Theories of changes in land use extent and intensity can be mapped in a two-dimensional space,
251 with places and actors ranked according to their degree of integration in markets on the one hand
252 and to their reliance on labor versus capital inputs on the other (Figure 1). Beyond describing land
253 productivity and its changes, intensification theories frequently predict how the efficiency of other
254 production factors change (Figure 2). These same theories can also explain land-use
255 disintensification and contraction, though legacies and path dependence may challenge their
256 application. For example, the accumulation of landesque capital – i.e., enduring anthropogenic
257 improvements in the productive capacity of land, such as through terraces and irrigation systems –
258 may hinder land abandonment even in the face of unfavorable production conditions (Håkansson
259 and Widgren 2016). We first discuss theories explaining changes in land use extent and intensity in
260 smallholder subsistence contexts, where households are units of decision-making, of production,
261 and of consumption, and directly interact with the environment (Section 3.2, lower-left quadrant of
262 Figure 1). We then consider theories of intensification and expansion dynamics when smallholders
263 progressively integrate into markets for inputs, outputs and consumer goods (Section 3.3, moving
264 right in Figure 1). We then move to land rent theories that are based on the neo-classical economic
265 framework under market conditions (Section 3.4). Section 3.5 incorporates the role of broader
266 institutions and social relations to develop theories of large-scale processes such as frontier
267 development. We end by introducing theories that incorporate non-linear land system dynamics and
268 feedbacks between the human and environmental components of land systems, as a bridge towards
269 Sections 4 and 5 (Section 3.6).

270

271 3.2 Theories on smallholder subsistence land use

272 Some theories describe the behaviors of smallholders or peasants farming for subsistence and
273 relying on labor as their primary input. These theories assume that smallholders pursue a satisficing
274 strategy aimed at maximizing labor productivity and avoiding the drudgery of labor. In the simplest
275 theory, referred to as “full belly”, a households' objective is to reach a certain subsistence target,

276 with minimal labor input (Kaimowitz and Angelsen 1998, Angelsen 1999). In the peasant theory of
277 Chayanov (1966), a household's labor inputs depend on the trade-off between addressing
278 consumption needs, which depend on household size, and the desire for leisure (time away from the
279 drudgery of farm labor), with no or little surplus produced for markets. This theory does not
280 explicitly discuss whether expansion or intensification is preferred to meet growing consumption
281 needs.

282 In frontier situations, where land and natural resources are abundant but labor and capital are scarce,
283 land-use expansion is expected to best render the satisficing outcome and thus be more likely
284 (Barbier 2010, le Polain de Waroux et al. 2018). In Boserup's theory (1965), intensification arises in
285 response to population pressure (i.e., higher ratio of population per suitable land available for
286 expansion; Erb et al. 2016b). This theory assumes that the technologies required for intensification
287 are available to farmers and explains what causes them to adopt these technologies. Intensification
288 is chosen over expansion only when land becomes scarce, because in non-mechanized systems the
289 marginal productivity gains of labor intensification are postulated to be decreasing—i.e.,
290 intensification raises land productivity but decreases labor productivity (Figure 2A, 2B). Output per
291 area and capita are only maintained if land productivity rises faster than labor productivity declines
292 (Figure 2A). Geertz (1963) used the term of “agricultural involution” to describe situations where
293 land productivity stagnates while labor productivity still declines. This involution path may
294 continue up to the point where output per population and area decreases, at which a Malthusian
295 crisis would occur, unless technological or institutional transformation induces a regime shift,
296 returning the system to a path of increasing land productivity (Figure 2A, 2B, Turner and Ali 1996).
297 Netting (1993) built on these theories to emphasize the specificity of smallholder households as
298 being both production and consumption units, allowing for flexible and low-cost family labor
299 inputs. Netting also showed that labor-based, agroecological forms of intensification have higher
300 energy efficiency (energy return per unit of energy invested, EROI) than capital-based forms of
301 intensification (Figure 2B).

302

303 3.3. Theories of induced intensification and institutional innovation with market integration

304 Induced intensification theory (Turner & Ali, 1996) extends Boserup's theory by acknowledging
305 that, firstly, demand per unit area constitutes a necessary but insufficient cause of intensification
306 because it is moderated by technological, institutional, and socioeconomic variables. Institutional
307 constraints on land-use expansion (e.g., land-use policies, tenure, or access rules) also influence
308 land accessibility and intensification possibilities beyond the physical availability of suitable land
309 (Section 3.5). Induced intensification theory also accounts for the role of biophysical attributes in

310 production. The most extremely prime and marginal lands tend to exacerbate intensification, owing
311 to their strong response to inputs on the one hand, or their considerable investments requirements,
312 which tend to be concentrated on small areas and induce path-dependent reinforcement of
313 intensification on the other hand.

314 Secondly, in addition to subsistence demand linked to population pressure, when land users engage
315 in markets, the demand for agricultural products also comes from other consumers. Households
316 may respond to these demands by separating subsistence from market cultivation, leading to
317 different levels of intensification. These responses depend on the degree of market engagement.
318 Pure subsistence systems are increasingly uncommon, but many land-use agents, particularly in
319 developing countries, face conditions of incomplete or imperfect market integration (de Janvry et al.
320 1991, Turner et al. 1993). Such conditions are explained by dominant positions of other market
321 actors, risks and transaction costs, and cultural norms, values, and practices (Laney and Turner
322 2015). Subsistence and commercial sectors can interact through multiple channels, and
323 smallholders' land-use decisions depend not only on their integration with cash crops markets but
324 also with other markets, such as those for buying staple food crops, for labor, and consumer goods
325 and accessing credit (de Janvry et al. 1991, Meyfroidt 2017). In particular, smallholders with
326 limited access to consumer good markets lack incentives to increase production to gain more
327 income, and are thus unlikely to respond to increased opportunities for marketing their surplus
328 production. Subsistence producers, which may not be directly in contact with agricultural products
329 markets, can still be affected by changes in market conditions indirectly through their effects on
330 land and labor demand (Dyer et al. 2012).

331 The related induced institutional innovation theory (Ruttan and Hayami 1984, Ruttan 1997) embeds
332 the long-term intensification processes within a broader market and institutional environment.
333 Technological innovation, not just its adoption, is an endogenous response to changing scarcity of
334 production factors, as institutions (governments, agribusiness companies) invest in developing
335 innovations that enhance labor or land productivity, determined by the scarcity of either. Over the
336 long term, this process is expected to increase the total factor productivity (TFP), which is the ratio
337 of total output compared to all inputs compounded (Coelli and Rao 2005, Fuglie 2015). TFP reflects
338 knowledge, skills, and technological shifts that enhance the productivity of land, labor, and capital
339 (Figure 2B).

340

341 3.4 (Neo)classical economic theories of land rent

342 (Neo)classical economic theories of land rent formalize expansion and intensification processes by
343 building on the underlying value, or rent, of the land, assuming that land will be used for the

344 activity that generates the largest expected value. In Ricardian theory, rent is a function of land's
345 biophysical characteristics, e.g., soil quality and water availability, and of the scarcity of land with
346 high productivity (Ricardo 1817). In contrast, von Thünen's location theory addresses the spatial
347 organization of land use surrounding a central market (von Thünen 1966). In this theory, land rent is
348 a function of the distance to this central market, which affects transportation costs depending on
349 perishability and bulkiness of the farm goods. This generates patterns where crops with high-value
350 and high transport costs are produced near the market, and less valuable and more easily
351 transportable ones are produced further away, holding farm production costs constant across space.
352 Furthermore, within each land use type, intensity of production declines with increasing distance to
353 market.

354 Land rent manifests itself through the bid rent, or, the maximum amount that any land user would
355 be willing to pay for using that land (Alonso 1964, Peet 1969). Land use extent and intensity change
356 along with bid rent changes, affected by a myriad of factors such as road building, new
357 technologies, climatic change, or market conditions. Such changes move the land-use frontier,
358 usually involving the expansion of the more profitable land use (Walker 2004, Angelsen 2010).
359 Originally, these theories described land use under market conditions where land can be bought or
360 rented, and where goods produced on land are sold on one local, central market. These assumptions
361 are generally relaxed when studying contemporary contexts. Local land uses often responded to
362 distant markets during the colonial era, generating intensive land uses a continent away from the
363 market in question (Peet 1969; Wallerstein 2011). Where land, labor, or products markets are
364 missing or incomplete, concepts of "shadow rents" or "shadow prices" are used (i.e., the value that
365 households put on marginal changes in these variables, Mundlak et al. 2004, Dyer et al. 2006).

366 Land rent theories underlie many land use simulation models (Irwin 2010, Parker et al. 2012,
367 Filatova 2015), and have been used extensively to explain agricultural change, deforestation
368 (Angelsen 2007, Walker 2004), and urban expansion (Alonso 1964, Sinclair 1967). In urban
369 contexts, further developments include the incorporation of environmental externalities and other
370 amenities into land rent theories (Bockstael and Irwin 2000, Clark et al. 2002), and theories on the
371 role of regulatory institutions such as markets and urban planning policies in shaping different types
372 of land rents (Jager 2003). Higher-level theories of urban land use change posit that cities are
373 "growth machines" organized to intensify land use and thus generate higher land rents (Molotch
374 1976). Land rent theories may explain broad land use patterns but may fail to explain specific local-
375 level (i.e., parcel) land-use change stemming from individual decision-making (Irwin and
376 Geoghegan 2001).

377

378 3.5 Institutional, political ecology and other theories of resource management, access, and
379 appropriation

380 Another set of theories explains the role of institutions, power, and agent heterogeneity in natural
381 resources management, and in large-scale dynamics of expansion, such as frontiers development or
382 intensification. These theories, drawing from political economy, political ecology, new institutional
383 economics and other sources, focus on the institutions and processes that determine how agents
384 access and use land and other resources needed for agricultural expansion and intensification (e.g.,
385 water for irrigation).

386 The theory of access proposes that technology, capital, markets, knowledge, authority, social
387 identities, and social relations shape how access to land and other resources can be gained,
388 controlled, and maintained (Ribot and Peluso 2003). Access to different resources (land, labor) and
389 capitals (financial, social, cultural) in turn shape the “capabilities” and agency to access additional
390 livelihoods resources and to make land-use decisions (Bebbington 1999). The “environmental
391 entitlements” framework further posits that institutions, defined as regularized patterns of
392 behaviors, mediate the relations between these heterogeneous endowments (assets, capitals) and the
393 entitlements that can be derive from them (Leach et al. 1999, Garrett et al. 2017). Political ecology
394 theories explain that institutions, such as the political and economic systems of the colonial era,
395 shape conflicts over access to environmental resources (Blaikie 1985, Bryant 1998). Other theories
396 also explain how and why farmers diversify towards off-farm activities, with important implications
397 for land use in terms of availability of labor, financial capital, and other resources (Barrett et al.
398 2001, Rigg 2006, Batterbury 2001).

399 One prominent institutional theory explains the conditions—called “design principles”—under
400 which different forms of self-governance arrangements of common-pool natural resources can lead
401 to sustainable outcomes. These include clearly defined boundaries of the resource, regulations
402 adapted to local conditions, collective decision-making processes that encompass most resource
403 appropriators, effective and accountable monitoring, graduated sanctions, mechanisms of conflict
404 resolution, and centralized governments that allow local institutions to self-organize (Ostrom 1990,
405 2005). Institutional theories also discuss the interactions between formal institutions and social
406 organizations (Bebbington 1996).

407 Theories integrating different insights have been proposed to explain the development of *frontiers*—
408 situations of resource appropriation where land and natural resources are abundant while labor and
409 capital are scarce. The resource frontiers theory expects that, where land is accessible to many
410 potential land users, and with population pressure and increasing affluence, rapid land-use
411 expansion occurs as land-use agents engage in a race for the accumulation of natural resources

412 (Barbier 2010). Frontier expansion has been described as moving from a populist or pioneer stage,
413 dominated by smallholders whose in-migration is often supported by state policies and
414 infrastructure, toward a capitalized or consolidated stage, where powerful actors consolidate land
415 into large holdings (Pacheco 2005, le Polain de Waroux et al. 2018). In the absence of state
416 planning, these corporate actors produce neoliberal frontiers (Hecht 2005; Brannstrom 2009), where
417 export-oriented farming is motivated more by global demand and deregulated access to land than by
418 government subsidies. The transition from populist to corporate frontiers is associated with two
419 processes. First, declining yields and profitability of pioneer agriculture (due to soil degradation or
420 poor soil quality) with out-migration of smallholders leads to the formation of a “hollow frontier”,
421 with depopulation and extensive land uses (Casetti and Gauthier 1977; Hecht 2005; Rudel et al.
422 2002). Second, a “technology treadmill” occurs when continuing competition leads to
423 intensification and the exclusion of farmers—mostly smallholders—that lag behind due to lack of
424 capital, technology, or knowledge, giving rise to large-scale capitalized agriculture (Levins and
425 Cochrane 1996; Chatalova et al. 2016). Smallholders might sell their land or be displaced, engage in
426 commercial operations as laborers, migrate to cities in a process of “de-agrarianization” (Bryceson
427 and Jamal 1997), or seek cheaper land elsewhere, driving further frontier expansion (Richards 2012,
428 2015, Section 4.2). Such treadmill occurs in various contexts, not only frontiers. “Commodity
429 frontiers” correspond to contexts dominated by large-scale commodity agriculture in which
430 “abnormal” rents (i.e., land rents much higher than land prices) caused by changes in technology,
431 regulations or other conditions, are maintained through imperfect land market conditions and the
432 heterogeneous capabilities of agents in appropriating these rents. The latter depends on these agents’
433 access to production factors and their information, preferences, and agency (le Polain de Waroux et
434 al. 2018).

435 Institutional and political ecology theories are also increasingly invoked to explain urban land
436 dynamics at different scales (Roy Chowdhury et al. 2011). Different theories explain that decisions
437 on urban landscapes, from private residential greenspace management to large-scale patterns of
438 urban expansion and densification (i.e., intensification), derive from nested factors at distinct social-
439 hierarchical scales. These factors include (i) household-level characteristics and environmental
440 attitudes (Larsen and Harlan 2006), and social stratification and lifestyle groups (e.g. Ecology of
441 Prestige – Grove et al. 2006); (ii) neighborhood-scale formal and informal institutions (Robbins and
442 Sharp 2003; Heynen et al. 2006) and housing filtering, i.e., the change in status and income of the
443 population of a neighborhood over time (Muth 1969; Bond and Coulson 1989); and (iii) municipal-
444 scale land-use governance such as planning and zoning (Munroe et al. 2005; Irwin and Bockstael
445 2007) and regional development trends in industrial and transportation infrastructure, and
446 differential diffusion (Geyer and Kontuly 1993). As with other land uses, urban areas can expand

447 but also disintensify and contract. Contemporary processes of recession, de-growth and de- (or
448 post-)industrialization, as well as shifting patterns of urban-rural connectivity and changing public
449 discourses are increasingly theorized to shape urban urban shrinkage and out-migration (Nelle et al.
450 2017). New, post-industrial economies and governance destabilize industrial urban morphologies by
451 introducing “creative” islands and edges of urban heritage, innovative building design, and new
452 urban open/public spaces and peripheries (Gospodini 2006). Recession landscapes also reflect
453 declining values of residents’ willingness to pay for environmental amenities, which affect patterns
454 of urban disintensification (Cho et al. 2011).

455

456 3.6 Theories of social-ecological feedbacks

457 Another set of theories focus on how feedbacks between human and ecological dynamics shape
458 land use. We highlight four influential approaches among a wide variety: regime shifts, resilience
459 dynamics, social-ecological systems, and farmers’ adaptation to environmental change.

460 Regime shifts are large, abrupt, persistent changes in the structure and function of ecosystems
461 (Biggs et al. 2012, Kull et al. 2017). Regime shifts highlight that land systems can experience
462 surprising, non-linear shifts from being around one set of mutually reinforcing structures and
463 processes to another, through the interaction of ‘fast’, such as weather or market fluctuations, and
464 ‘slow’ processes, such as erosion of crop diversity (Scheffer et al. 2001, Biggs et al. 2012, Müller et
465 al. 2014, Ramankutty and Coomes 2016). Such theories have focused on economic, bio-cultural,
466 biophysical and health feedbacks mechanisms that can produce ‘poverty traps’, e.g., through
467 agricultural involution (Lade et al. 2017), as well as on how human-mediated ecological processes
468 such as fire, nutrient cycling, grazing, and water flows regulate regime shifts (Gordon et al. 2008,
469 Bestelmeyer et al. 2015, Jepsen et al. 2015, Rocha et al. 2015). In land systems, the regime shifts
470 perspective helped understand ecological dynamics such as savanna-forest and tundra-forest
471 transitions as influenced by fire and climatic conditions, and herbivory pressure (Rocha et al. 2015).
472 This perspective underlies theories of land-use transitions (Section 5).

473 Another stream of research focusing on regional environmental management and governance
474 highlights diversity and disturbance as key aspects of the resilience of social-ecological systems
475 (Berkes et al. 2008). Successful intensification usually simplifies ecosystems, reducing their self-
476 regulatory capacity (Holling and Meffe 1996). Societal dependence on land products necessitates
477 increasing investments in artificial regulation to stabilize outputs, which may lead to a ‘rigidity trap’
478 in which a large part of the output is absorbed for maintaining production, or to a transition to
479 another land system with new actors and land uses (Allison and Hobbes 2004, Vang Rasmussen and

480 Reenberg 2012, Goulden et al. 2013). Research on path dependence and sunk costs in land systems,
481 in relation to landesque capital, builds on this perspective (Janssen et al. 2003).
482 Works building on the social-ecological systems perspective explored how formal and informal
483 institutions shape and are shaped by social-ecological interactions (Ostrom 2009), and how features
484 of society and ecosystem create fit and misfit between social and ecological dynamics, and enable
485 or impair collective action to address shared environmental problems (Janssen et al. 2007). These
486 studies highlight that institutions regulating natural resources are diverse and include governments,
487 other public institutions, traditional regulations, and cultural norms.
488 Finally, a large stream of literature proposes theories on how environmental signals are incorporated
489 into decision-making (Verburg 2006, Lambin and Meyfroidt 2010, Meyfroidt 2013). These insights
490 suggest that land use intensification or changing regulations can arise as a response to the
491 degradation of ecosystem services linked to expansion into natural ecosystems under the conditions
492 that this degradation is perceived, interpreted and valued. These perceptions and valuation of
493 environmental change build on cultural backgrounds of what constitute valuable ecosystem services
494 (Daniel et al. 2012). Multiple works have also explored the conditions under which farmers are
495 expected to adopt innovative agricultural practices in response to climate and other environmental
496 change (Prokopy et al. 2008, Niles et al. 2015).

497
498

499 4. Theories of land-use spillovers

500 We here propose middle-range theories of complex land system processes, such as land-use
501 spillovers and displacement, combining several of the theories discussed above. *Land-use*
502 *displacement* refers to the separation between places of production and consumption, but has been
503 used in a broader sense to refer to geographic shifts of land use from one place to another
504 (Meyfroidt et al. 2013). *Land-use spillovers*, which can explain some forms of displacement, refer
505 to situations where land-use changes or direct interventions on land use (e.g., policy, program, new
506 technologies) in one place have impacts on land use in another place. With globalization and
507 increasing complexity in land-use change processes, land-use spillovers have constituted a focus for
508 research over the recent years (Lambin and Meyfroidt 2011). Theoretical synthesis on the various
509 forms of land-use spillovers and the mechanisms and conditions under which they occur is thus
510 timely. Various forms of spillovers have been distinguished, including leakage, indirect land-use
511 change, and rebound effects. *Leakage* is a form of spillover caused by a land-use intervention, such
512 as an environmental conservation policy, which triggers land-use change elsewhere that reduces the
513 overall benefit of the local intervention (Meyfroidt and Lambin 2009, Ostwald and Henders 2014).

514 *Indirect land-use change (iLUC)* is a land-use change in one place caused by a land-use change in
515 another place (Lapola et al. 2010). Following this definition, all land-use leakage occurs through
516 iLUC. A *rebound effect* is a form of spillover where adoption of intensifying practices stimulates
517 land-use expansion (Angelsen and Kaimowitz 1999, 2001, Lambin and Meyfroidt 2011). Other
518 spillovers have been shown, e.g., between agriculture and forestry through the increased
519 consumption of wood pallets to export agricultural products (Jadin et al. 2016a).

520

521 4.1 Land sparing and rebound effect

522 Intensification is often promoted to fulfill growing societal demands for land-based products, while
523 reducing pressure on land and thus preserving nature, an effect called land sparing or the Borlaug
524 hypothesis (Figure 3). Land sparing can be *absolute*, i.e., resulting in net farmland contraction
525 (Rudel et al. 2009). Globally, intensification of staple crops through the Green Revolution has
526 resulted in *relative* land sparing, i.e., reducing per-capita land demand or the rate of agricultural
527 expansion compared to the counterfactual scenario without intensification, although net agricultural
528 area still increased (Stevenson et al. 2014). The potential impacts of land sparing on biodiversity
529 conservation and livelihoods are debated (Loos et al. 2014, Kremen 2015, Phalan et al. 2016,
530 Fischer et al. 2017). Key assumptions are that intensification spares land and that this land is
531 returned to nature. In reality, intensification can also lead to a rebound effect, i.e., a form of
532 spillover where adoption of intensifying practices stimulates land-use expansion (Angelsen and
533 Kaimowitz 1999, 2001, Lambin and Meyfroidt 2011). Such rebound effect, also known as Jevons'
534 paradox, occurs when intensification increases the profitability of agriculture. Theories identify the
535 conditions under which intensification can spare land – potentially allowing for nature conservation
536 – or is more likely to lead to a rebound effect. Rebound effects include the direct response of the
537 original agents reinvesting an additional income or spared production factors into expansion. It also
538 includes indirect, systemic effects linking increased land-use efficiency to land-use expansion such
539 as through increased consumption of other goods thanks to lower spending on goods produced more
540 efficiently (Greening et al. 2000, Sorrell et al. 2007).

541 At local scales, land sparing is more likely to occur when intensification increases local production
542 costs per unit, i.e., when intensification requires scarce and thus expensive capital or labor inputs
543 such as irrigated paddy fields (Villoria et al. 2014, Byerlee et al. 2014) (Figure 3, Place A), and
544 when there are strong biophysical, institutional, or other restrictions on accessing land, or high
545 demand for environmental amenities (Rudel et al. 2009, Meyfroidt and Lambin 2011, Phalan et al.
546 2016). Local land sparing is also more likely when the demand for the product is inelastic to price,
547 i.e., when although intensification makes agriculture more efficient and less costly, this decrease in

548 production costs does not lead to an increase in demand, as in the case of inferior goods (such as
549 staple caloric crops), or when markets are closed (Hertel et al. 2014). In contrast, a rebound effect is
550 more likely to occur at local level when there are low physical or institutional restrictions on land-
551 use expansion and when demand is elastic to price, as is the case for superior goods such as meat,
552 luxury or leisure crops (e.g., cocoa, coffee), feed or bioenergy crops (Lambin and Meyfroidt 2011),
553 or when markets are well-integrated and the intensifying region or producer is large enough to
554 influence prices (e.g., soybean in South America) (Hertel et al. 2014). A rebound effect is also likely
555 when intensification occurs by switching to highly profitable production alternatives with high
556 income elasticity of demand, i.e., products for which demand increases strongly when income rises,
557 such as meat in developing countries. Finally, a local rebound effect is likely when intensification
558 results in increased competitiveness, i.e., the ability and performance of a producer or a region to
559 sell products in a given market compared to the ability of competing regions and producers.
560 Competitiveness increases with intensification when initial yields were low in the intensifying
561 region, and when production costs decrease, as manifested by TFP increases, such as with
562 agglomeration economies (Section 4.3) or low cost of capital and labor (Hertel et al. 2014).

563 The local land sparing or rebound effect may further affect regional or global land use through
564 markets, depending on the level of market integration (Figure 3, Place B). When local land-sparing
565 intensification is associated with increased local production costs, it may trigger an upward effect
566 on prices when the affected region is large (Hertel et al. 2014, upper part of Figure 3). If the demand
567 on this larger market is elastic to price, demand will reduce, without or with few impacts on land
568 use elsewhere. But if the demand is inelastic to prices and thus stable, the upward shock on price
569 may trigger an intensification or expansion in places that, relatively, gain competitive advantage,
570 and an acceleration of land sparing in the initial place that has lost competitiveness. In the medium
571 or long term, local intensification may promote economic growth and wage increases, and thus
572 cause an indirect rebound effect, locally or globally, by stimulating consumption. Conversely, a
573 rebound effect that makes local land use more competitive may, when happening in a large region,
574 trigger a downward shock on prices in broader markets (Angelsen and Kaimowitz 2001, Villoria et
575 al. 2014, lower part of Figure 3). If demand is inelastic to prices, it may induce land sparing in other
576 regions that have lost competitiveness. When demand is elastic to prices or is increasing because it
577 is elastic to income and wages are rising, the downward effect on prices triggers increased demand
578 and thus further intensification or expansion. When the local and distant effects on land use go in
579 opposite directions, the net balance in area of land being used depends (i) on the relative yields of
580 the place where the initial intensification occurs, versus those where the expansion occurs (Hertel et
581 al. 2014), and (ii) on whether changes in production in distant places occur through changes in area

582 or intensity. An additional condition for absolute land sparing is that productivity increases faster
583 than the demand.

584

585 4.2 Leakage and indirect land-use change

586 Policies aimed at setting aside land or restricting land-use expansion can result in leakage through
587 different pathways, each of which can be explained by a specific middle-range theory (Figure 4).
588 An effective policy may reduce availability of land directly through restricting access or indirectly
589 through set-aside incentives. In turn, increasing the scarcity of land would result in increasing land
590 price, and possibly decreasing the profitability of land use. We identify four leakage mechanisms,
591 which interact in reality: *activity* leakage, *land-market* leakage, *commodity-market* leakage, and
592 *supply-chain* leakage. Leakage is a challenge for example in policies aiming at reducing
593 deforestation, such as through protected areas, the Reducing Emissions from Deforestation and
594 Forest Degradation (REDD+) scheme, or the New York Declaration on Forest and other multi-
595 stakeholders zero-deforestation agreements.

596 *Activity leakage* occurs when production factors or inputs are highly mobile such that labor and
597 capital used on the land targeted by the restrictions are reallocated to places with available and
598 accessible land (Atmadja and Verchot 2012, Lim et al. 2017, Pfaff and Robalino 2017). This
599 pathway is akin to the “pollution haven hypothesis” (le Polain de Waroux et al. 2016), according to
600 which polluting companies react to environmental legislation by moving their activities to places
601 with fewer restrictions. Activity leakage is more likely to occur through labor reallocation under
602 conditions of subsistence agriculture, with lack of off-farm alternatives or cultural preferences for
603 land-based activities, and through capital reallocation when sunk costs of capital investments in the
604 initial place are not too large (e.g., extensive cattle ranching which has little fixed assets) (Atmadja
605 and Verchot 2012, Henders and Ostwald 2012). Unfavorable conditions for intensification locally,
606 and growing demand for the affected product reinforce this pathway by creating incentives for
607 producers to continue production elsewhere.

608 *Land-market leakage* can also occur, where appreciation of land rent in the affected place spreads
609 through land markets to land situated elsewhere, driving land investments, including deforestation,
610 in these places (Richards 2015). In the region affected by regulations, increase in price of the non-
611 affected land can facilitate activity leakage by providing landowners with financial capital to
612 reinvest elsewhere, as suggested for Brazilian Amazon frontiers (Arima et al. 2018, Richards and
613 Arima 2018). Although in principle this path can occur as leakage from policy restrictions, it is
614 more likely to occur as a form of indirect land use change resulting from an increase in demand of a

615 commodity, as exemplified by soybean demand increasing land prices and driving cattle expansion
616 in the Brazilian Amazon (Arima et al. 2018, Richards and Arima 2018).

617 The initial intervention can result in a decrease in production when land is set aside, regulations are
618 imposed on input use, production costs are increased, or expansion is restricted (le Polain de
619 Waroux et al. 2017, Lim et al. 2017, Pfaff and Robalino, 2016). A large shock on production relative
620 to the size of the product's market—when the affected region is large (Hertel et al. 2014) or in
621 smaller, segmented markets—may trigger a price increase, depending on the degree of market
622 integration and the price transmission for this good. This market shock can be absorbed in three
623 ways. Firstly, the greater the demand-elasticity to price, the more the price increase is absorbed on
624 the demand side through a reduction in consumption of the affected good. Secondly, intensification
625 is likely to occur, locally or distantly, if labor or capital inputs, TFP-enhancing technologies, and
626 production technologies with flexible input-ratios –i.e., which can accommodate diverse
627 combinations of labor and capital-based inputs– are available, and if land supply is restricted
628 (Wunder 2008, Börner et al. 2017). A similar result could be obtained if the affected good can be
629 substituted by another one with lower land demand. Thirdly, if the intervention takes place in a
630 high-yielding region and the conditions for intensification are not met in the different regions where
631 production takes place, the market shock is more likely to be absorbed through *commodity-market*
632 *leakage*, in which land use expands elsewhere in response to changes in product prices. Examples
633 are the possible restructuring of timber markets in response to policies for Reducing Emissions from
634 Deforestation and Forest Degradation (REDD+) in tropical regions (Jonsson et al. 2012) and of
635 soybean markets under changing environmental governance (le Polain de Waroux et al. 2017). Such
636 leakage may also occur if the affected good is substituted by more land-demanding goods.

637 The same chain of spillovers can also be triggered when the existing land use is replaced by another
638 land use. The effect is then called indirect land-use change. This pathway can be due to a policy
639 supporting the good derived from the second land use (e.g., a biofuel policy or supply chain
640 intervention that support the production of a “clean” alternative good), or any other process leading
641 to an expansion or increase in demand for that second good. Processes of large-scale land
642 acquisitions (also referred to as “land grabs”) and technology treadmill (Section 3.5) can also lead
643 to indirect land-use change, in particular through the *activity* mechanism.

644 Finally, a supply-chain intervention that excludes a given good or suppliers who do not meet
645 sustainability standards can also result in leakage along the pathways described above (Alix-Garcia
646 and Gibbs, 2017, le Polain de Waroux et al. 2017). But in addition, it can also lead to a *supply-chain*
647 *leakage*, where producers continue to produce the same good but shift to other buyers, sell their
648 products by “laundering” them through intermediaries that are compliant with the intervention, or

649 switch to producing another good with high environmental impacts (Rausch and Gibbs 2016,
650 Lambin et al. 2018). We propose that this leakage occurs if (i) the incentives to improve production
651 practices are insufficient; (ii) the origin of a product can be easily concealed due to complexity in
652 the life cycle of production, (iii) the stickiness or rigidity in supply chains is low—due to low
653 transaction cost or substitutable goods—, i.e., producers, buyers, and intermediaries can easily shift
654 their activities from one market to another (Villoria and Hertel 2011); and (iv) the affected
655 producers have a competitive advantage against producers elsewhere or can switch to alternative,
656 profitable land uses.

657

658 4.3 Other processes of land-use spillovers and displacement

659 Other spillover effects such as specialization of regions through clustering of specific activities have
660 been explored with economic geography theories of competitive advantage due to technology
661 spillovers (Porter 2000, Fujita et al., 1999). These theories formalize an economic tradeoff between
662 centripetal forces promoting spatial concentration of economic activities, and centrifugal forces
663 triggering their dispersion. The agglomeration factors for a particular activity in each location (i.e.,
664 travel costs to markets and jobs, availability of skilled labor, innovation spillovers, and social
665 amenities) are weighted against dispersion factors (i.e., density, land prices, negative spatial
666 environmental externalities). From a business strategy perspective, clustering and intensification of
667 related activities in specific locations lead to path-dependent increasing returns to scale when output
668 increases faster than inputs (Krugman, 1991). The influence of agglomeration economies, or
669 positive externalities associated with the clustering of activities, has been widely used in urban
670 studies to explain how the migration of workers to a city eventually gives rise to an increase in
671 goods and services available, which drives further migration (Fujita and Krugman 1995). The
672 concept of agglomeration economies has also been increasingly integrated into land-use studies in
673 the agricultural and forestry sectors to understand regional variations in land rents that are not
674 explained by Ricardian and Thunian theories (Garrett et al. 2013, Richards 2017).

675

676 5. Theories of land-use transitions

677 Changes in land-use extent and intensity interact to produce non-linear trajectories of land systems.
678 Rapid, non-linear changes in land resource uses are driven by positive feedbacks, where initial
679 interventions or disturbances precipitate a cascade of further changes (Peters et al. 2004,
680 Ramankutty & Coomes, 2016). These dynamics produce land-use transitions, which are structural
681 transformations of land systems from one dynamic equilibrium to another (Lambin and Meyfroidt

682 2010, Müller et al. 2014), akin to regime shifts in complex systems theory (Scheffer et al. 2001,
683 Biggs et al. 2012, Filatova et al. 2016, Kull et al. 2017, Section 3.6). The development of theories of
684 land-use transition constitute a key achievement of recent land system science, which we synthesize
685 here.

686

687 5.1 Forest transition theories

688 One well-studied type of land-use transition is forest transition, which describes a structural shift
689 from net forest loss to net forest gain through natural regeneration or planted forests (Mather 1992).
690 Forest transitions in the 19th and early 20th centuries occurred mainly in temperate, developed
691 regions, but are increasingly observed in tropical regions as well (Meyfroidt and Lambin 2011).

692 Three land-use dynamics explain how land is made available for restoration of natural ecosystems
693 and reforestation in one social-ecological system, apart from a decrease in demand (Jadin et al.
694 2016b) (Figure 5). Firstly, agricultural and forestry intensification can lead to abandonment and
695 reforestation (Green et al. 2005, Section 4.1). Secondly, a spatial redistribution of land use to better
696 match land suitability in increasingly integrated markets may also result in intensification and land
697 sparing (Mather & Needle, 1998, Nanni and Grau, 2017). Thirdly, international trade in land-based
698 products may facilitate forest recovery in one place by displacing pressure on environments
699 elsewhere, as leakage or in response to changes in global markets (Meyfroidt et al. 2010, Section
700 4.2). Reforestation can arise from natural regeneration on abandoned land, or from tree plantation or
701 assisted nature restoration.

702 Processes of forest transition have been described through interacting pathways combining these
703 three land-use dynamics and influenced by multiple drivers (Rudel et al., 2005, Meyfroidt and
704 Lambin 2011, Liu et al., 2017; de Jong et al., 2017,). Each of these pathways has been explained by
705 one middle-range theory of forest transition (Figure 5). The *economic development* theory
706 highlights urbanization and industrialization driving labor scarcity in agriculture, and intensifying
707 and concentrating production on the most suitable land, thereby retiring marginal agricultural lands
708 from production (Rudel et al., 2005). Substitution of wood-fuel by fossil and other energy carriers
709 also strongly contributes (Erb et al. 2008). This mechanism is often framed at the national level,
710 ignoring cross-border leakage or international labor migration. The influence of these international
711 processes on national-scale reforestation have been integrated in the *globalization* theory
712 (Mansfield et al., 2010; Meyfroidt and Lambin, 2009; Kull et al., 2007; Hecht and Saatchi, 2007; Li
713 et al., 2017; Jadin et al., 2016b). Trade may correspond to deforestation leakage to countries with
714 less strict environmental regulations (Section 4.2). International trade may also facilitate the
715 concentration of land use on the most suitable lands, possibly relieving pressure on marginal

716 ecosystems in a global-scale land sparing (Section 4.1) (Kastner et al., 2014, Youn et al., 2017). The
717 diffusion of global efforts and ideologies for nature conservation, such as biodiversity protection or
718 carbon sequestration, may also drive reforestation (Hecht and Saatchi, 2007; Kull et al., 2007).

719 In the *forest scarcity* theory, economic, political, and cultural responses to environmental
720 degradation and the scarcity of forest products and services drive forestry intensification, tree
721 plantation and rehabilitation, land set-asides, and protection of remaining natural habitats (Hyde,
722 Amacher, & Magrath, 1996; Lambin & Meyfroidt, 2010; Rudel et al., 2005, Park and Youn, 2017).
723 This is a form of regime shift characterized by negative feedbacks. Government actions to protect
724 and plant forests may follow various motives beyond forest scarcity, including geopolitics, state
725 consolidation, and prejudices against minority groups (Peluso and Vandergeest, 2001; Hecht et al.,
726 2014, McElwee 2016). These state actions fall under a *state forest policy* theory, which can be
727 considered a variant of the forest scarcity one (Lambin & Meyfroidt, 2010). Governmental support
728 for tree plantation often lies at the intersection of economic development and forest policies (Zhang
729 et al., 2017; Cochard et al., 2017).

730 The *smallholder, tree-based intensification* theory results from dynamics at the smallholder farm
731 scale that influence planting or maintenance of trees (Rudel et al., 2002; Pokorny and de Jong,
732 2015). Dynamics of agroforestry, sylvo-pastoral management and gardens can result in a “tree cover
733 transition” that extends beyond forests in the strict sense (van Noordwijk et al. 2014). Other
734 pathways have been suggested, including the “impacts of war and conflict”, which ascribe either
735 deforestation or forest recovery to side effects of geopolitical conflict (Hecht et al., 2014, Hecht and
736 Saatchi, 2007; Robert-Charmeteau, 2015; de Bremond, 2013). These theories of forest transition
737 have been formalized through several disciplinary lenses, such as land rent frameworks (Barbier et
738 al., 2010). The different pathways of forest transition lead to distinct ecological impacts and
739 environmental values of returning forests (Kull 2017, Wilson et al. 2017).

740

741 5.2 Other land-use transition theories

742 Paralleling forest transition theories, urban theories explain cycles of urban growth, decline and
743 renewal (Clark et al. 2002). Stylized theories of land-use transition, more akin to grand theories,
744 posit that land use in a region follows a series of transitions that accompany socioeconomic
745 development and changes in societal metabolism. Sequences run from wildlands with low human
746 population densities dependent on hunting, foraging, resource extraction, and extensive use of fire,
747 to frontier clearing for subsistence agriculture, and increasingly intensive and commercial
748 agricultural systems, ultimately leading to intensive industrial agriculture supporting large urban
749 populations, and the abandonment of low-suitability agricultural lands (DeFries, Foley, & Asner,

750 2004; Fischer-Kowalski & Haberl, 2007, Jepsen et al. 2015). These theories aim to explain long-
751 term land-use trajectories, which are presented as a directional modernization process akin to
752 Rostow's stages of growth (Rostow 1960). It also evokes the environmental Kuznets curve, which
753 posits increasing environmental degradation in early stages of economic development and a reversal
754 with higher income, in a trajectory moderated by policies (Barbier, Burgess, & Grainger, 2010).
755 These transitions can be theorized as resource substitution and problem shifting, where adoption of
756 intensive fossil fuel-based land use displaces impacts from land systems towards climate (Erb et al.
757 2008). In a similar way, sociocultural niche construction theory explains long-term changes in
758 human societal scale and transformation of the biosphere through land use as the product of
759 sociocultural evolution in subsistence regimes based on ecosystem engineering, social
760 specialization, and non-kin exchange (Ellis 2015). Some of these stylized theories have been
761 criticized for being overly deterministic, simplifying the actual complexities of land-use trajectories,
762 and ignoring trade, geopolitics, and other relations between regions (Perz, 2007; Walker, 2008). Yet,
763 they provide a bird's eye perspective and umbrella frameworks under which more specific middle-
764 range theories can be formulated.

765

766 6. Directions for further theory development

767 Middle-range theories of land system change can be formulated to synthesize key processes of land-
768 use change and the conditions under which these processes manifest. Such middle-range theories
769 provide a constructive path towards more generalized knowledge of human-environment systems
770 (Magliocca et al. 2018). These theories remain to be further tested and refined, particularly
771 regarding the conditions leading to different pathways. Several emerging trends in land systems
772 require further theoretical development, including relations between urban and rural areas (Seto et
773 al. 2012), such as central flow theory (Taylor et al. 2010); transnational land acquisition (or "land
774 grabbing") and land speculation, and associated land-use displacement and conflicts (Zomers 2010,
775 McMichael 2012); and new forms of private and hybrid land governance where supply chain actors
776 and consumers promote the adoption of voluntary sustainability standards and sourcing practices
777 (Rueda et al. 2017, Lambin et al. 2018, Lambin and Thorlakson 2018).

778 Connectivity among distant places is increasing, and globalization plays an increasingly important
779 role in driving land systems dynamics through trade but also via information flows and increased
780 human mobility. Beyond understanding the conditions and mechanisms under which spillovers
781 occur, further theoretical developments are required to explain how connectivity—access to land,
782 markets, technology, information, and financial capital—institutions, and sustainability values
783 shape the susceptibility of different places to receive spillovers. A stronger integration with social

784 theory would allow conceptualizing spillovers as a land use moving across a social as well as
785 physical space, and investigating interactions between these two spaces (Faust et al. 1999).

786 Theoretical progress will continue to arise from integration with other fields that bring different
787 perspectives on land systems. This encompasses explanations of the emergence of certain spatial
788 patterns of land uses that shape landscape structure and influence social and environmental impacts
789 (Middendorp et al. 2016). Land-use theories incorporate socio-economic and institutional drivers of
790 land system change, but could better account for interactions between biophysical processes and
791 human behaviors at multiple scales (Lade et al. 2017), and the short- and long-term co-evolutionary
792 dynamics between human societies and environments (Gual and Norgaard 2010; Ellis 2015; Waring
793 et al. 2015). Theories could also improve understanding of cross-scale interplays between macro-
794 level phenomena, such as landscape patterns, structures or new functions, and micro-level
795 interactions by building on landscape ecology (Wu, 2013), complex adaptive systems (Page 2015),
796 and sociological approaches on micro-macro interactions (Coleman, 1990). Research on common-
797 pool resources (fisheries, water systems) can also provide insights on cooperation and collective
798 action for sustainable land management. Similarly, research on ecosystem services captures
799 multiple connections between people and nature across landscapes. Theories of land-use spillovers
800 and displacement could better integrate various streams of theories that focus on globalization and
801 telecoupling (Liu et al. 2013, Friis et al. 2016), including: (i) critical sociological theories related to
802 world system theories, such as dependency theory (Frank 1978), core-periphery relations, and
803 (ecological) unequal exchange (Hornborg 1998, Muradian and Martiez-Alier 2001, Moran et al.
804 2013), (ii) sociological and economic geography theories that analyze the structure and functioning
805 of global production networks and global value chains, and financialization of land uses (Gereffi et
806 al. 2005, Munroe et al. 2014, Isakson 2014), and (iii) anthropological theories of “scapes” that
807 articulate the different types of information and cultural flows that link distant places (Escobar
808 2001, Tsing 2005, Niewöhner et al., 2016). Spillover theories could also better integrate
809 environmental processes such as Earth system teleconnections from land-use changes to distant
810 regions via changes in rainfall regimes (Keys et al. 2017), spillovers between biodiversity and
811 ecosystem services (Maestre Andrés et al. 2012), or poverty traps linked to land degradation and
812 loss of resilience (Barrett and Bevis 2015, Lade et al. 2017, Haider et al. 2018).

813 Land system science theories largely rely on the rational actor and expected utility theory as a basis
814 for apprehending human decision-making and behavior. The theories hence assume that agents act
815 purposefully to attain their goals and maximize their well-being according to their expectations
816 (Meyfroidt 2013, Groeneveld et al. 2017). Agent behavior can take the simple form of selfish utility
817 maximization with perfect information, or more refined forms of bounded rationality models

818 (Simon 1956, Gotts et al. 2003), other-regarding preferences (Fehr and Fischbacher 2002, Sautter et
819 al. 2011), or prospect theory (Ligmann-Zielinska 2009). Land system theories could benefit from
820 incorporating theories of human behavior to address the diversity of motivations and cognitive
821 processes as well as the role of social networks that determine land-use choices, responses to
822 environmental change, and social norms related to land use (Meyfroidt 2013, Schlüter et al. 2017,
823 Groeneveld et al. 2017, van Duinen et al. 2015). For example, empirical studies on smallholder
824 agriculture or ecosystem services are now considering diverse forms of individual and collective
825 relationships to nature that include sense of place and embodiment (Masterson et al. 2017,
826 Raymond et al. 2017), psychological theories of planned behavior (Schwarz and Ernst 2009), and
827 theories from behavioral economics (Nyborg et al. 2016) that account for decisions that deviate
828 from the expected utility theory.

829 We barely addressed the normative aspects of sustainability and the tradeoffs and synergies between
830 social, economic, and environmental objectives (DeFries et al. 2004, DeFries and Nagendra 2017).
831 Developing interventions to improve the sustainability of land systems first requires solid
832 theoretical foundations on the causes of land-use and land-cover change and how they can be
833 influenced. The discourses on “sustainable intensification”, for example, often rely insufficiently on
834 theoretical knowledge of intensification dynamics and their spillovers and trade-offs (Loos et al.
835 2014, Rockström et al. 2017).

836

837 7. Conclusion

838 A major challenge in land system theories is that land is simultaneously a biophysical entity, a
839 territory, a commodity, a habitat for nonhuman species, a resource for productive activities, and a
840 buffer for absorbing pollutants. It is allocated, regulated, and administrated by various laws, norms,
841 and rules. It is also a source of meaning and sense of place, a landscape component, and
842 symbolically loaded. Theories of the causes of land system changes cross theoretically and
843 epistemologically disparate knowledge domains, and build on deductive, abductive, and inductive
844 approaches. A grand, integrated theory of land system changes remains elusive. Nevertheless, the
845 past decades have seen the elucidation of chains of causal mechanisms that explain well-bounded
846 phenomena, and the conditions and contexts under which they occur, laying the foundations for
847 middle-range theories on how, why, when, and where land systems change. We have reviewed the
848 major theories of land-use expansion and contraction, intensification and disintensification, and
849 synthesized theories on land-use spillovers including leakage, indirect land-use change, rebound
850 effect, and land-use transitions, with a focus on agriculture and forestry. Middle-range theories
851 come in a nested way, with different degrees of generality: Different pathways leading to a given

852 process each correspond to a theory, but our consolidated account of each of these land system
853 processes is a higher-level middle-range theory. A similar approach articulating chains of causal
854 mechanisms, and the conditions and contexts under which they operate, could be applied to other
855 land and social-ecological systems processes to enrich the portfolio of middle-range theories.
856 Theories of change for sustainability governance would be strengthened by building on causal
857 chains derived from such middle-range theories. Along with basic frameworks, case-specific
858 explanations, and grand theoretical schemes, middle-range theory development constitutes an
859 important endeavor for land system science and for the study of human-environment interactions
860 and sustainability science.

861

862

863 Figures captions

864

865

866 **Figure 1. Main theories of land-use expansion and intensification**

867 Theories are mapped by the contexts and agents for which they have been formulated,
868 though they can be used for other contexts. The X-axis distinguishes contexts and agents
869 according to their degree of market integration and commercialization of land use. This axis
870 encompasses the (i) accessibility of markets for inputs (e.g., financial capital, skilled labor,
871 machinery, agrochemicals, but also land); (ii) accessibility and reliability of markets for
872 outputs; and, (iii) share of the farm output which is marketed, or importance of markets for
873 livelihoods. Not all these dimensions necessarily coincide (e.g., large companies in frontiers
874 situations can be fully integrated into global outputs markets but face imperfect land
875 markets; le Polain de Waroux et al. 2018). The Y-axis distinguishes land-use agents
876 (households, farms, companies) in their degree of reliance on labor versus capital-based
877 inputs. Most land uses, e.g., cropland, grazing lands, forestry, agroforestry systems, can fall
878 under various degrees of inputs types and market orientation, depending on the specific
879 context and agents, with different land users operating in the same landscape possibly having
880 different positions in this graph.

881

882 **Figure 2. Theoretical trajectories of land-use intensification and changes in productivity**

883 Fig. 2A. Classic trajectories of land-use intensification articulating the theories of Boserup,
884 Malthus, and Geertz (adapted from Ellis et al. 2013, inspired by Turner and Ali 1996). Fig.
885 2B. Intensification trajectories expressed in terms of labor, energetic and economic
886 efficiency versus land productivity. Labor productivity is expected to decline under
887 Boserupian intensification without mechanization or higher capital intensity (e.g., more
888 fertilizers applications per unit area) (Section 3.2). Energetic efficiency (measured as energy
889 return per unit of energy invested, EROI) is expected to decline under most forms of
890 intensification, though it declines more strongly under capital-based, industrial
891 intensification than under agroecological intensification. With decreasing labor productivity
892 under Boserupian intensification, or lower energy efficiency under industrial intensification,
893 total factor productivity (TFP) may decline unless new technologies or institutional
894 arrangements allow a structural transformation of the land-use system towards higher TFP,
895 until reaching again a point of decreasing marginal returns and possible decline (inspired by
896 Stone (2001).

897

898 **Figure 3. Theory of land sparing and rebound effect with intensification**

899 The causal chain starts from the left with an intensification event, which can be induced by
900 pressure on land, new technologies, or other factors (Section 3). Each arrow describes a
901 causal link between two events, under the conditions indicated in the boxes overlaid on the
902 arrows. The left panel (Place A, dark green) describes the effects on the place where this
903 initial intensification occurs. The right panel (Place B, light green) describes potential effects
904 on the broader market (or local market if the initial place is disconnected from broader
905 markets), and possible feedbacks to the initial place.

906

907 **Figure 4. Theory of leakage and indirect land-use change (iLUC)**

908 The causal chain starts from the left with an intervention restricting land use in one place
909 (Place A, dark green box). Each arrow describes a causal link between two events, under the
910 conditions indicated in the boxes overlaid on the arrows. The leakage pathways on the right
911 panel (Place B, light green box) can take place locally or distantly, depending on conditions
912 that make local or distant places more susceptible to receive leakage. The three triggers on
913 the top-left give rise to the same land-use processes, but these are then called iLUC. Leakage
914 pathways can also be triggered by a supply chain intervention that bans a given good or
915 production method (e.g., if its production entails deforestation).

916

917 **Figure 5. Theories of forest transition**

918 The central panel, starting from the left, describes the processes within a given area (e.g.,
919 country) for which forest transition is observed. Each arrow describes a causal link between
920 two events, under the conditions indicated in the boxes overlaid on the arrows. The upper
921 and lower horizontal panels describe exogenous drivers linked to globalization. Background
922 colors indicate the different pathways (i.e., middle-range theories) of forest transition.

923

924

925 References

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1497 **Figures**

1498

1499 **Figure 1. Main theories of land-use expansion and intensification**

1500

1501 **Figure 2. Theoretical trajectories of land-use intensification and changes in productivity**

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1503 **Figure 3. Theory of land sparing and rebound effect with intensification**

1504

1505 **Figure 4. Theory of leakage and indirect land-use change (iLUC)**

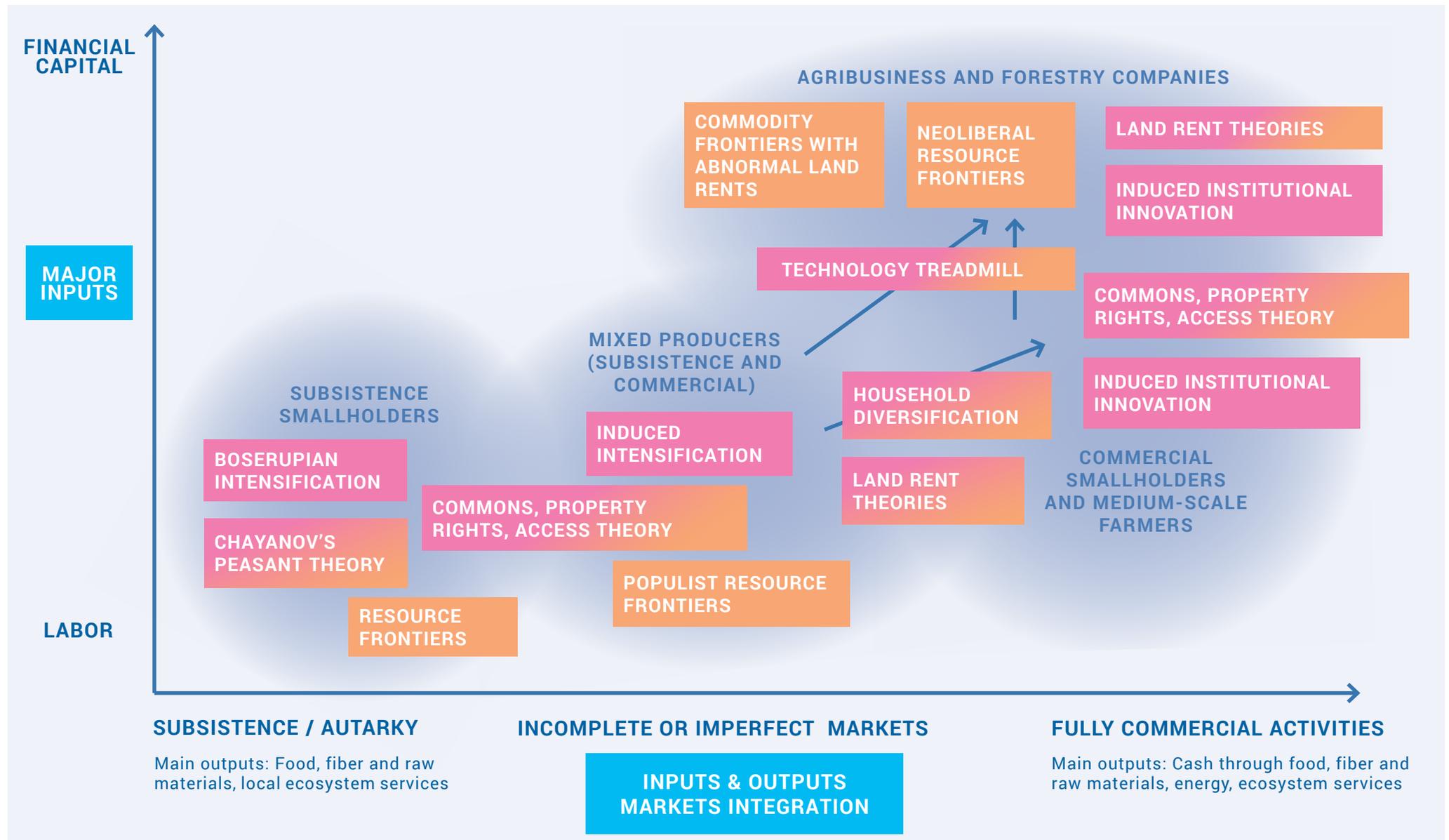
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1507 **Figure 5. Theories of forest transition**

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FIGURE 1 | Main theories of land use expansion and intensification



Land use agents
Dimensions

Theories focused on expansion
Theories focused on intensification

→ Theories that focus on how land systems and actors move across the graph

FIGURE 2 | Theoretical trajectories of land-use intensification and changes in productivity

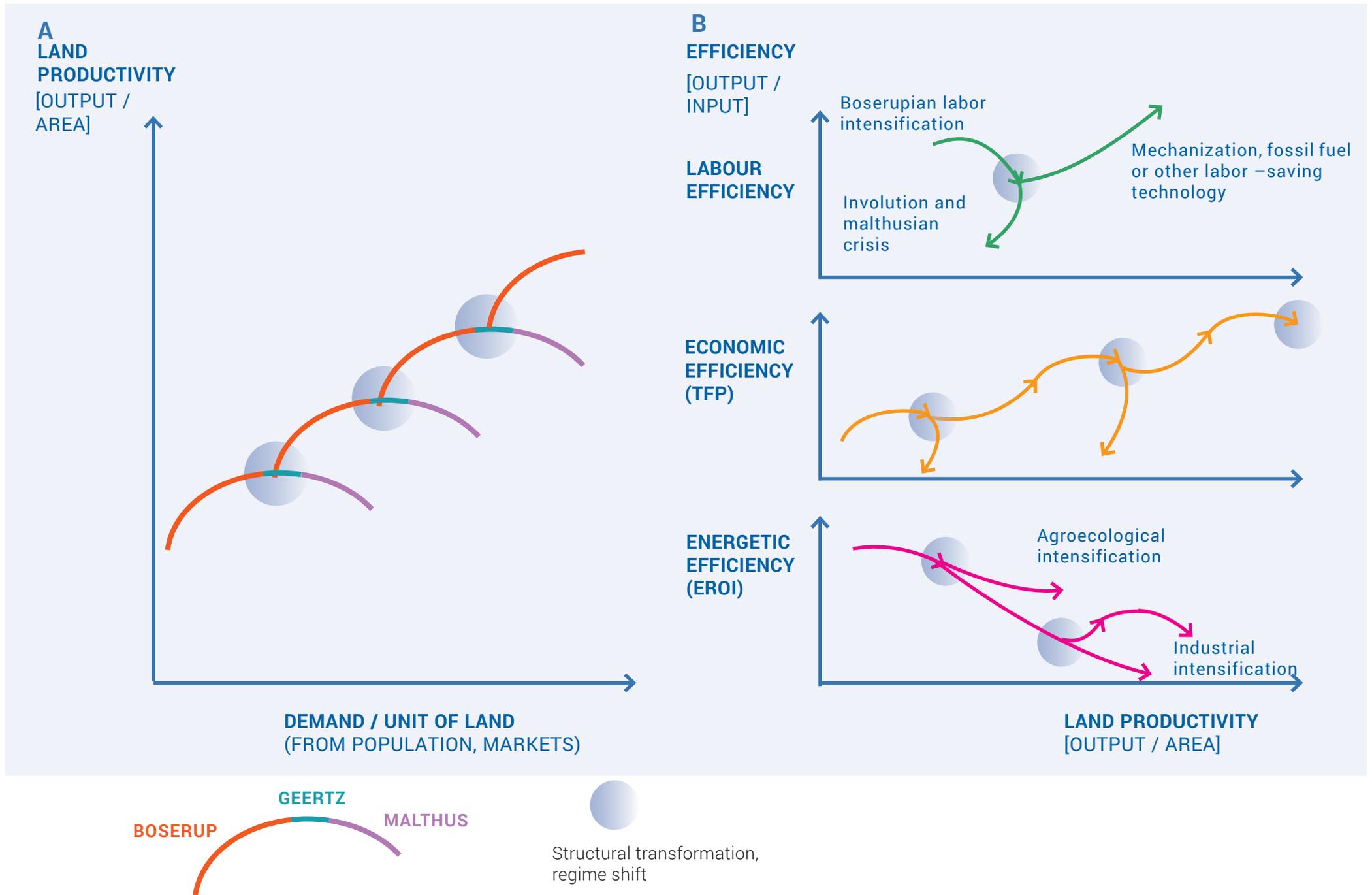
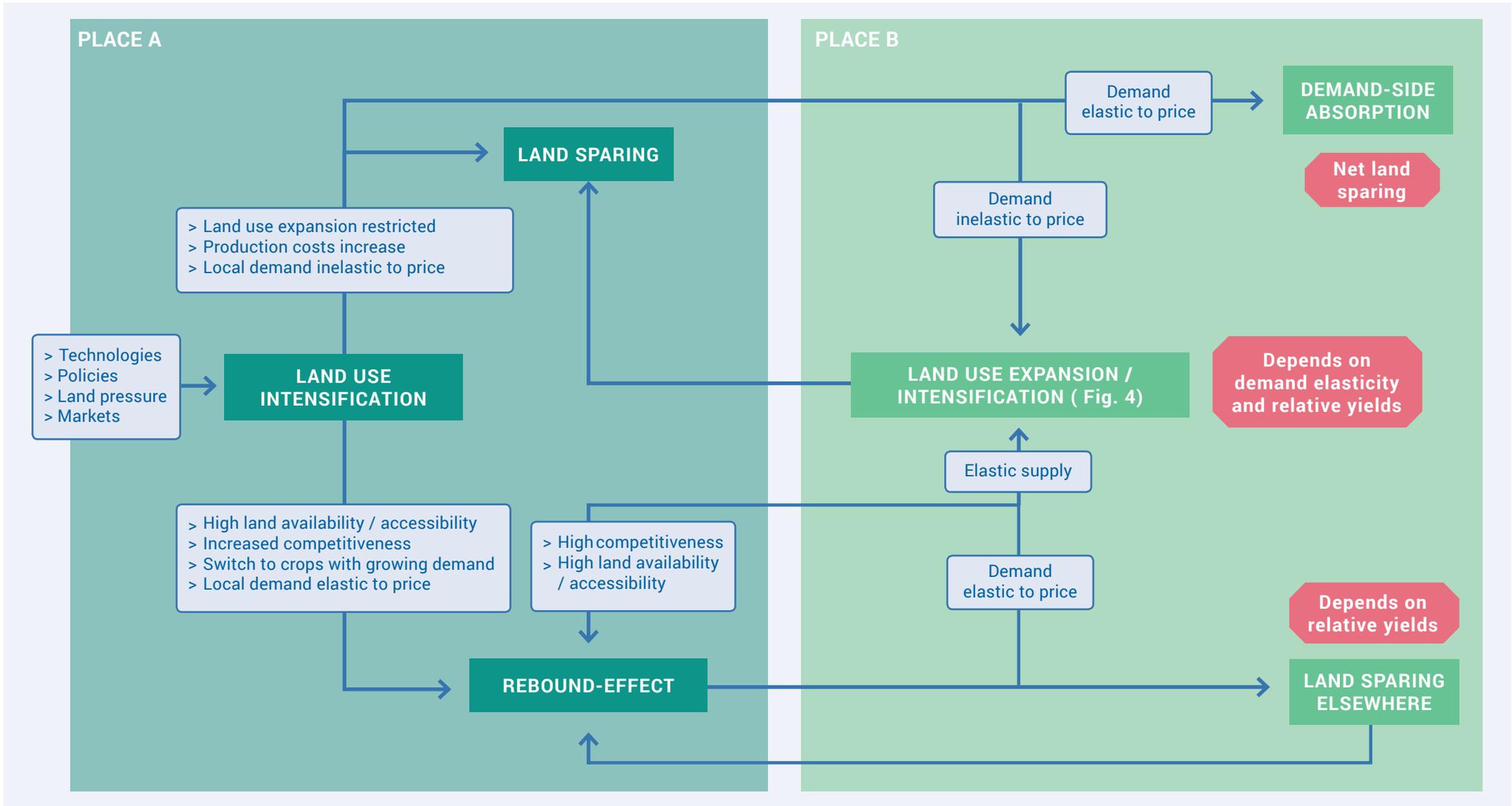
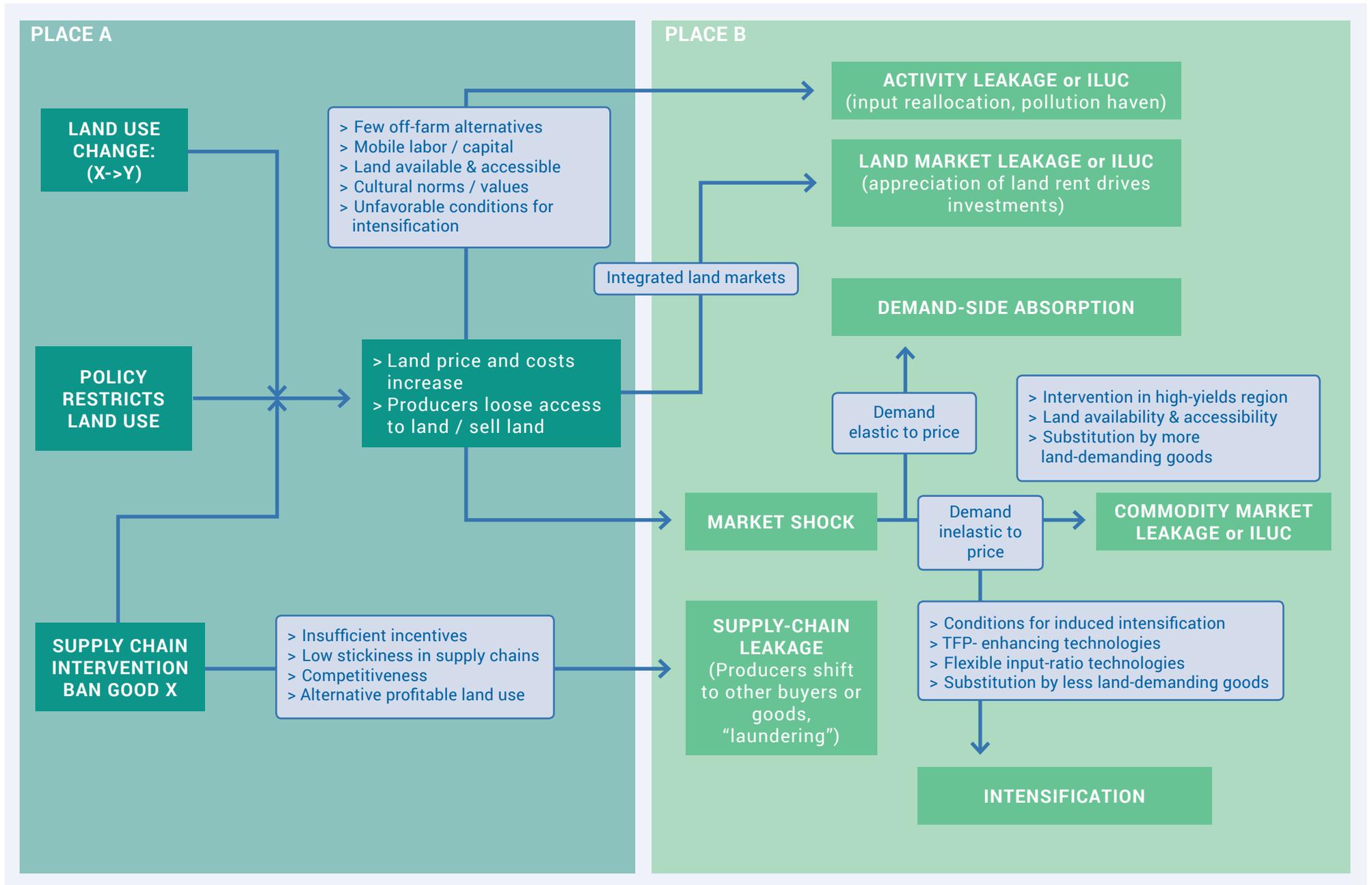


FIGURE 3 | Theory of land sparing and rebound effect with intensification



- Place A
- Place B / distant effects up to global scale (depends on market integration)
- Conditions
- Net (global) outcome

FIGURE 4 | Theories of leakage and indirect land use change (iLUC)



Place A
 Land use spillovers, local or distant (up to global) (depends on market integration)
 Conditions

